Photodegradation of Real Pharmaceutical Wastewater with Titanium Dioxide, Zinc Oxide, and Hydrogen Peroxide During UV Treatment

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Abstract:-This paper presents the photocatalytic degradation of real pharmaceutical wastewater from Abbot Laboratories (Private) Limited, Karachi, Pakistan, using TiO₂, ZnO, and H₂O₂. The pretreated sample wastewater was used for degradation experiments and tests were carried out at 38 0 C under pH of 9 and 4 in a stirrer bath reactor equipped with eight ultraviolet tubes. The use of this technique is common in removal of the organic, inorganic pollutants and pathogens. Optimal conditions were selected from reported results of the researchers on advanced oxidation processes (AOPs) for removal of residual pharmaceuticals from real pharmaceutical wastewater. The three catalysts (Titanium dioxide, zinc oxide, and TiO₂/ H₂O₂) used in this study are effective catalysts in photocatalytic degradation of real pharmaceutical wastewater. The maximum degradation achieved was 45.11% by combined use of TiO₂ and H₂O₂ at 38 0 C and pH. The degradation improved at higher pH with Zinc oxide and Titanium oxide. The results indicate that for real pharmaceutical wastewater, combined use of TiO₂/ H₂O₂ is comparatively more effective than ZnO and TiO₂ alone. The degradation of the pharmaceutical wastewater followed pseudo-first-order kinetics. The reaction rate constant was 0.00148/min for TiO₂/ H₂O₂. This study demonstrates that for real pharmaceutical wastewater reacts differently to catalyst than synthetic pharmaceutical wastewater, or formulated wastewater.

Keywords:-*Hydrogen peroxide, Pharmaceutical wastewater, Photocatalytic degradation, Titanium dioxide, Ultra violet light, Zinc oxide*

I.

INTRODUCTION

The pharmaceuticals are found in drug manufacturing wastewater, hospital wastewater, and industrial wastewater of pharmaceutical origin. The pharmaceutical usually consist of biologically active substances designed as lipophilic and are resistant to biodegradation. In aquatic media these substances if discharged to the environment, have potential for accumulation and persistence in the environment. Though they appear at low concentrations, but still may impose serious effects on the environment. During last two decades researchers have focused their attention in searching suitable technologies to destroy the xenobiotic substances and recently been able to report advanced oxidation processes (AOPs) as highly efficient treatment process for pharmaceutical wastewater.

Among many processes the ozone oxidation is adapted as part of the treatment at several water production plants. Treatment of pharmaceutical wastewater by AOPs is expensive, hence photocatalysis is an option considered widely by researchers. In recent years photocatalytic degradation of pharmaceutical wastewater has received attention of the scientific community. The results of past research carried out on real pharmaceutical effluent are comparable to this study; however, the results of studies on aqueous solution, synthetic wastewater, formulation effluent, synthetic wash-water from antibiotics packaging distilled water, industrial wastewater of pharmaceutical origin, hospital wastewater, synthetic wash-water from medium scale drug manufacturing plant, biologically pre-heated pharmaceutical wastewater, actual wash-water from ointment manufacturing plant, effluent from bulk drug manufacturing plant, municipal wastewater, and secondary treated domestic effluent are worth considering. The previous findings on aqueous system indicate degradation of wastewater by TiO₂ and UV alone is negligible [1] and that coupling of TiO₂ with UV greatly enhance degradation rates [2-4]. Only UV processes require long irradiation time [5] and only TiO₂ require high catalyst loading [6]. The variables affecting the degradation rates are catalyst loading, pH and temperature of solution/wastewater, irradiation time and intensity, and use of additional catalyst like H₂O₂ with and without Fe^{2+} . The particle size and surface area/_{BET} of the catalyst (TiO₂) is an important parameter too [3]. The total organic carbon (TOC) removal rate depends on refractory character of the organic-inorganic wastewater content.

Fatemeh reports high degradation of analgesics in water at pH 12 and 50 °C [7] whereas, Liming found degradation between pH of 3.5 and 9.5 [1]. The synthetic wastewater made from human and veterinary antibiotics show that degradation increases with increasing pH (pH range 7-12), H₂O₂ enhance performance [8,9]. The results of Kaniou et al on Sulfamethzine in distilled water indicate higher removal performance of H₂O₂ and ZnO was more active than TiO₂ [10]. 90% removal of Diclofenac from distilled water is reported by TiO₂ after 60 minutes of irradiation [11]. 55% COD removal from industrial wastewater of pharmaceutical origin is reported by San Sebastian by using H₂O₂ at pH 4 [12]. Xing et al found 94% color and 73% COD removal with $Fe^{2+}/H_2O_2at pH 5$ from biologically pre-heated pharmaceutical wastewater [13]. Kulik et al reports 87%-96% COD removal from actual wash water from ointment manufacturing plant with Fe^{2+}/H_2O_2 [14]. Augmentation of TiO₂ by H_2O_2 is reported by many researchers [15-18] and now it is believed that presence of H₂O₂ enhances degradation performance of the catalyst [9]. The generation of the OH radicals in the UV/H_2O_2 occurs via reaction (1) [19]. $H_2O_2 + hv \rightarrow 2^{\circ}OH$ (1)

The photogenerated electrons can recombine and make their ways to the surface of TiO_2 , where they can react with species adsorbed onto the catalyst surface.

$$hv^{+} + H_{2}O \rightarrow OH + H^{+}$$

$$hv^{+} + OH \rightarrow OH$$

$$Organics + hv^{+} \rightarrow products$$

$$(2)$$

$$(3)$$

$$(4)$$

Organics + $hv^+ \rightarrow$ products

The photogenerated electrons can also combine with O₂ and organics and produce: $e^{-} + O_2 \rightarrow O_2^{-}$ (5) Organics + e^{-} > products (6)

The produced hydroxyl radicals (OH) and superoxide radical anion (O_2^{-}) can mineralize organics to water and CO₂ (end products): Organics + ((OH), O_2^{-} , etc.) \rightarrow

(7)Only a few studies have been conducted on industrial wastewater of pharmaceutical origin $employing TiO_{2and}H_2O_{2as}$ the catalysts with varying degradation results. It is still unclear that how realpharmaceuticalwastewaterwouldrespon activity to photocatalytic of widely used catalysts.ThispaperreportsthefindingsoftheresearchworkemployingTiO₂,ZnO,and H₂O₂and the performance of photocatalytic degradation reactor of real for а pharmaceutical wastewater. The results of this study would aid to the application of nanophoto catalystparticles forremovalofcontaminantsfrompharmaceuticalwastewater.

II. **EXPERIMENTAL PROCEDURE**

2.1 material

Inthisstudy titanium dioxide powder (anatase) formandZincoxide (Dentam) were used as supplied.TheseChemicalsaresuppliedbyBDHwithpurityof99.99% with the average particle $size of 25 \pm$ m^2 $4nmandspecificsurface of 55\pm$ g⁻¹.The30% AqueousSolutionofHydrogen 8 peroxide was also supplied by BDH. The pH of the sample in the reaction vessel was adjustedwithcalculatedvolume/weightof1NHClor1 NNaOH.Thestandardsolutionandreagentswere usedformeasuring theCODinclosed reflux colorimeterasproposed **byAPHA** [20].These catalystsareknownforoptimalcatalyticefficiencyandgoodinterparticlecontactsinwater[21].

2.2 ultraviolet energy

The selectionofultraviolet(UV)energydependsontheabsorptioncapacityofthemoleculesof thecontaminantin thesamplewater. The UV energy not only advances the oxidation process but micro-organism. Thephotodegradation sterilizesthewastewaterbydestroyingfivemajorgroupof efficiencydependson UVdosage[22].Previousresearch[10,15,23]hasindicatedthat UVlightat wavelength254-579 nm, and 50-150 Wis adequate forsynthetic and real pharmaceutical wastewater with an irradiation time of 1ofmostdegradationis 4hour.However,occurrence reported within first hour of treatment. Hence is there as on forse lecting of the parameters for this study.

2.3 experimental setup

selected for this study was 8.0 in cheslong with an UV output of 11 W. A totalThe UV lamp ApparatusCo., numberofeightUVlamptubes, manufacturedbyJiangsuShenXingPhotoelectricity inches,volume5 liter)to Ltd.China, we resubmersed in the specially designed reaction vessel (diameter 8 produce88Wofenergy.Thepretreatedsamplewastewaterwasplacedina 5 literglass photoreactioncell. The catalystpowder was suspended in the cylinderverticalreactoractingasa sampleandthesolutionin thecellwaskepthomogenousbyconstantstirringwitha topmounted stirrerandaerationbyusinganairpumpasshownintheFig. 1 (actualexperimentalrig). The schematic diagram of the

experimental setupis shown in Fig. 2. The aerial view of the geometrical arrangement of the UV tubes is shown in Fig. 3. The passing of the air facilitated

the wastewater circulation around the reaction vessel in order to maintain the reaction temperature at the desired value. This cell was placed in a constant temperature bath. The pH meter and temperature probe was inserted into the cell to monit or pH and temperature of the wastewater.

 $The photocatalytic degradation was carried out over suspension of Titanium Oxide, Zinc Oxide and TiO_{2/H_{2}O_{2}} under ultraviolet irradiation.$



Fig. 1 photocatalytic oxidation reactor



schematic diagram of the experimental set-up

Fig. 2



Fig. 3 the aerial view of the geometrical arrangement of the uv tubes in the reactor

TheSpectronic "Genesys 20" Spectrophotometer with wavelength range 325 to 1100 nm, accuracy \pm 2.0nm wasused to measure absorbance at λ max of 625 nm. The rate of photodegradation iscalculatedinpercentagewithtimeofirradiation. Testswere conducted for

variables such as pH of sample was tewater, type of catalyst, and exposure time of UV light.

2.4 procedure

 $\begin{array}{cccc} realpharmaceutical wastewater which is placed in a photoreaction cell. For experiments with hydrogen peroxide, \\ 2.4 & ml & H_2O_230\% & purity & was added & to sample & wastewater. & The \\ suspension was subjected to irradiation under UV light for 2 & hour. The suspension was stirred and \\ a erated throughout the experiment. & & & \\ \end{array}$

 $\label{eq:action} A tregular time intervals, the 2 cm^3 of irradiated sample was taken out from the reaction vessel with the help of microsyring e, and then filtered through Millipore filter of 0.45 um$

 $solid catalyst. Using cuvettes the absorbance of the supernatant liquid is measured at \lambda max of 625 \ nm.$

The antibacterial activities of the wastewater before and after the photocatalysis were measured by Abbot Laboratories using HPLC to determine the remaining concentration of the parent antibiotics (Sulfonamide, Fluroquinolones, Bacteriostatics, and Penicillingroup); analgesics (Acetaminophen); parent anti-inflammatory drugs (Diclofenac and Ibuprofen); and CNS stimulant (Caffeine). TABLE 1 lists the characteristics of sampled pharmaceutical wastewater from Abbot Laboratories, and TABLE 2 shows the minimum standard discharge limits for pharmaceutical effluents [25].

Table 1	. Characteristics of	Sampled Pharmace	utical Wastewater fr	rom Abbott Laboratories	Limited
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Characteristics	Name of Drug	Value
COD (mgxl ⁻¹)		168-240
Antibiotics (mgxl ⁻¹)	Sulfonamide	0.02-0.52
	Fluoroquinolones	5-43
	Bacteriostatic	0.08-0.33
	Penicillin group	0.002-0.022
Analgesics/Antipyretics	Acetaminophen	10-28.55
Anti-inflammatory	Diclofenac	0.01-480
drugs		
	Ibuprofen	0.47-885
CNS Stimulant	Caffeine	2.8-10.56
TDS $(mgxl^{-1})$		980-1240
TSS (mgxl ⁻¹)		105-200
P ^H (dimensionless)		7.4-7.8
Temperature (^o C)		24-30

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Table 2. Winning Standard Disenarge Ennits for Tharmaceutical Enfuents [25]				
Parameter	Composition (mgxl ⁻¹)			
COD	250			
BOD (3 days, 27 ^o C)	10-40			
Oil & Grease	10			
TSS	100			
P ^H (dimensionless)	6.0-			
Mercury	0.01			
Arsenic	0.20			
Chromium (Cr ⁶⁺)	0.10			
Lead	0.10			
Cyanide	0.10			
Phenolics (C_6H_5OH)	1.00			
Chromium (Cr ⁶⁺)	0.10			
Sulphides (as S)	2.10			
Phosphates (as P)	5.0			

Table 2. Minimum Standard Discharge Limits for Pharmaceutical Effluents [25]
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Each culture tubes were incubated for 12 hours at 38 0 C. After the incubation, the bacterial growth was measured using the optical density at 625 nm. All bio-assessment tests were repeated for accuracy. The optical density value was converted to growth inhibition, I, and, was calculated in terms of changes in absorption spectra. The degradation efficiency I (%) was calculated as:

$I = [A]_o - [A]_t / [A]_o \ge 100$

Where $[A]_o$ is the initial value and $[A]_t$ after any irradiation time. All experiments were carried out at two pH of sample with different photo catalyst. For accuracy of results the absorbance at a given time was compared with a calibration curve. The calibration plot was obtained by using a known percentage of real pharmaceutical wastewater. The calibration of spectrophotometer was done using manual procedure outlined in Spectronic "Genesys 20" and solutions were prepared from BDH products.

III. RESULTS AND DISCUSSIONS

3.1 Effect of Catalyst Type

The results presented in Fig. 4 shows effect of catalyst as an oxidant on degradation ofreal pharmaceuticalwastewaterbyphotoxidationprocessatconstantpH9andFig.5showseffectof pH on degradation ofwastewater withTiO₂. The wastewater waseffectively degraded using titanium dioxide or zincoxide under pH of 9. The maximum degradation achieved within 120minutesof irradiationtime, wasabout 33.23% by using TiO2 and 40.23% by using ZnOat 38⁰C. There are limited research finding sontheuse of catalyst like ZnO. Most researchers have focusedontheuse of TiO₂,H₂O₂,Fe²⁺/H₂O₂,andFe²⁺/H₂O₂/UV.Theresultsof thisstudyarenotdirectly comparableto reportedfindingsdueto theparametersstudied.However,the variationsin effectivenessofthecatalyst canbediscussed duetosimilartrendinCODandTOCremoval efficiency. Kaniou reports that ZnO is more active than TiO₂[10]. Theyachieved degradationin60minuteswhenTiO₂wascombinedwithSiO₂.Whereas,Limingfoundmorethan

90% pollutant's (paracetamolinaqueoussolution) degradation in 80 minutes at pH9.5 by TiO_2 alone[1]. This trend is comparable to our findings. The effectiveness of ZnO washigher than TiO_2 at given pH and temp. Reduced effectiveness of TiO_2 is due to medium pH and lower reaction temperature. Fatemeh found higher degradation at 50 °C. At high pH formation of OH radicals is enhanced [7]. The higher reactivity of ZnO is attributed to the geometry and working conditions of the photo reactor. The saturation level of ZnO under given operating conditions is lower than TiO_2 .

(8)



Fig.

4. The effect of catalyst on degradation of pharmaceutical waste water at pH 9.0, T=38°C



Fig. 5. The effect of pH on degradation of waste water with TiO_2 (Temperature = 37°C)

3.2 Effect of pHwith additional catalyst H_2O_2

TheeffectofpHondegradationofpharmaceuticalwastewaterbyTiO₂/H₂O₂

isshowninFig.6.Themaximumdegradationachievedwithin120minutesofirradiationtime,atpH9wasabout 45.11%.ByaddingthecatalystH₂O₂to TiO2, the degradation was increased by 12%. It has been addition increases degradation efficiency of the catalystreportedbymanyresearchersthatH₂O₂ [22.26-30]. Yeon [30] foundmore than 99% removal of a moving from a queous solution in the presence of H₂O₂. Xiang-RongreportsthatFentonprocessismoreefficientthanUV/H₂O₂inthe absenceofTiO₂[6]asrateofreactionincreaseswithH₂O₂. Itisprovedthatmorethan90% degradationcanbeachievedinaqueoussolutionbutnotinrealpharmaceuticalwastewater. As real wastewater characteristics and chemistry of pollutants are significantly different than syntheticwastewateroraqueoussolution,it is expectedthatdegradationwouldvarywidely.Fig.7 and8 showperformanceofcatalystatconstantpH9 and4respectively.AtpH4 themaximum degradationachieved within 120 minutes of irradiation time, was about 40.18%. It indicated that $pHplayed an important role and the moderate photocatalysis in the presence of H_2O_2$ isdueto enhancedabsorbanceofUVlightcapacitybythecompound;whichissimilartothefindingsof manyotherresearchers[12,14, 31]. Even with Fentonoxidation process with H₂O₂, the maximum COD removal fromindustrialwastewater ofpharmaceutical origin was 55%. Hence itisnot surprisingtoachieveamaximumdegradationof45% usingreal pharmaceutical wastewater.A possibleexplanationforlowerdegradationcouldbe thatphotocatalyticreactioninrealwastewater

 $produces several intermediate products and these products do not transform into {\rm CO}_2 to transformed in a queous solution.$

theextent



Fig. 6 The effect of pH on degradation of wastewater by TiO_2/H_2O_2







Fig. 8 The performance of catalyst at constant pH of 9

3.3 Effect of reaction time

A linear relationship as shown in Fig. 9 was observed between irradiation time and COD removal. The results show that in oxidation process, the reaction rate plays an important role than adsorption rate. The OH⁻ radicals can degrade organic pollutants to intermediates and then the intermediates are further degraded to CO_2 and H_2O [7]. The reaction rate attains a maximum value at high pH; and the generation of OH⁻ radicals by the effect of irradiation on the TiO₂ of the composite may be another factor for increasing reaction rate in basic environment [1]. The maximum COD removal efficiency in 120 min was obtained at pH of 9, which corresponds with the findings of San Sebastian [12,22]. The comparability of reported work on treatment of pharmaceutical in water and wastewater by AOPs with current study is given in TABLE 3.



Fig. 9 The linear regression of degradation of pharmaceutical wastewater at pH 9 and temperature 38 $^{\rm o}$ C

	AOPfeature	<u>Scale</u>	<u>Measureof</u>	<u>Comparability</u>	Reference
Synthetic wastewater	O3aloneorO3/H2O2;	Bench	COD,TOC,	Yes	[8, 9, 16]
Municipal	Ozalone	Pilot	TOC	No	Notlisted
Formulation Effluent	O3alone,UValone, orO3/H2O2; Fe ²⁺ /H2O2	Bench	COD,TOC, BOD5	Yes	[9, 15, 16]
Syntheticwash- waterfromantibiotic packagingdistilled	O3alone,O3/MnSO4	Bench	COD,BOD5	No	Notlisted
Industrial wastewaterof pharmaceutical origin	UVonly,Fe ²⁺ /H ₂ O ₂ , UV/H ₂ O ₂	Bench	COD,BOD	Yes	[5, 12]
Hospitalwastewater	Fe ²⁺ /H2O2	Bench	COD,BOD5,	Yes	[31]
Syntheticwash- waterfrommedium scaledrug	Fe ²⁺ /H ₂ O ₂	Bench and Full	COD,BOD5	Yes	[17]
Biologicallypre- treated pharmaceutical wastewater	Fe ²⁺ /H ₂ O ₂	Bench	COD	Yes	[13]
Actualwash-water fromointment manufacturingplant	Fe ²⁺ /H ₂ O ₂	Bench	COD,BOD7	Yes	[14]
Effluentfrombulk drugmanufacturing plant	TiO2asanode, Graphiteascathode, Cl ⁻ aselectrolyte	Bench	COD,TOC	No	Notlisted
Distilled water/Deionized Water	O3alone,UValone, UV/Fe ²⁺ ,UV/Cu ²⁺ , TiO ₂ /UV,	Bench	TOC	No	Notlisted

Table 3. Reported Work on Treatment of Pharmaceutical in Water and Wastewater by AOPS

3.4 Kinetic analysis

The linear regression analysis of degradation of pharmaceutical wastewater with TiO_2/H_2O_2 at pH 9 is shown in figure 9. The plotted data produced a straight line that indicates that the degradation of the pharmaceutical wastewater can be described by the following first-order kinetic model equation.

 $\label{eq:constant} \begin{array}{ll} ln \ (C_o/C_t) = k_1 t \equation \\ (9) \\ Where C_o is the initial concentration value, kis there action rate, and tis the time. The equation \\ (9) presents kinetic expression and the correlation between ln (C_o/C_t) and reaction time t was \\ linear with R^2 = 0.9802 for TiO_2/H_2O_2. The value of determination coefficient R^2 clearly indicates \\ that this degradation process follows pseudo-first-order kinetics. The figure 9 also shows the rate \\ constant (0.0148/min). These results are very close to the findings of Sun Jian Hui [28]. Similar to \\ previous studies [13,21,32]. More precise rate of reaction can be calculated if effects of photon \\ absorption are included to the kinetic model equation. \\ \end{array}$

3.5 Pharmaceutical Wastewater and Mineralization Study

The composition of pharmaceutical wastewater is complex and may contain different types of organic pollutants. At the time of study, Abbot Laboratories were processing formulations for antibiotics, analgesics, anti-inflammatory drugs and CNS stimulant as listed in TABLE 1. Hence it is assumed that the same group of compounds would be present in wastewater. The HPL Cresults confirmed the concentration of the second pollutant as listed in TABLE 1.

During the degradation process, some intermediate products are formed which may be more difficult to degrade. The measurement of COD is related to total concentration of organic compounds in the waste and treated water. Hence the change in COD before and after treatment reflects degree of mineralization. The COD of sampled wastewater was in the range of 168 to 240 mgl/l, which is within the standard discharge limits. However, the composition of the treated water suggests that nearly all of the intermediates were mineralized by photocatalysis under given conditions. It is most likely that the treated water is free from toxic compounds. This study confirms that photocatalytic degradation can completely mineralize pharmaceutical wastewater but still lacks in comprehensive study that encircles the comparison of the effect of TiO₂, ZnO, and other catalysts under different operating parameters. Therefore, in order to commercialize the photocatalytic wastewater technology, several key parameters like catalyst development, reactor design and process optimization need to be further investigated.

IV. CONCLUSIONS

All three catalystsnamelyTitaniumdioxide,zincoxide,andTiO₂/H₂O₂usedinthisstudyare effectivecatalystsinphotocatalyticdegradation frealpharmaceuticalwastewater. However,the maximumdegradationachievedwas45.11% bycombineduseofTiO₂andH₂O₂at38^OCandpH of9, within 120 minutes of irradiations. With the same catalyst atpH of 4 the degradation decreased to 35.82%. The degradation at higher pH supports the notion that pH isone of the most

 $\label{eq:constraint} important operating parameter that affect the phot catalytic reactivity of the catalyst. The maximum degradation with ZnOwas 40.23\% and with TiO_2 was 33.23\% at pH9. The degradation of the pharmaceutical was tewater followed pseudo-first-order kinetics. The reaction rate constant was 0.0148/minfor TiO_2/H_2O_2. The results indicate that for real pharmaceutical was tewater, combined use of TiO_2/H_2O_2 is comparatively more effective than ZnO and TiO_2 alone. This study demonstrates that real pharmaceutical was tewater reacts differently to catalysts than synthetic pharmaceutical was tewater.$

 $The results of this study indicated that photocatalytic degradation can be {\constrained on the study} and {\constrained on$ usedforthereal pharmaceutical wastewater contaminated bydifferent inorganic-organic pollutants. Inorder to achieve more than 90% degradation, pretreated effluent is more suitable for AOP treatment. Due the second secondto low consumptionofchemicalsand energy, heterogeneous photocatalysis seems to be а promisingmethodundermoderntechnologywithbroadapplicationin treatingwastewaterform pharmaceuticalindustry, hospitals and municipalities. However, further studies are requiredfor identifying i) thebioactivity ofunknown intermediate products andby-products, ii)inhibition of photocatalyticactivitybynaturalorganicmatterpresentinthewastewater, and iii) comparison of theeffectofTiO₂ZnO,andothercatalystsunderdifferentoperatingparameters.

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